

1 Activity, energy use and species identity affect

2 eDNA shedding in freshwater fish

3

4 Bettina Thalinger^{1,2*}, Andreas Rieder¹, Anna Teuffenbach¹, Yannick Pütz¹, Thorsten
5 Schwerte¹, Josef Wanzenböck³ & Michael Traugott^{1,4}

6

7 ¹ Department of Zoology, University of Innsbruck, Technikerstr. 25, 6020 Innsbruck, Austria

8 ² Centre for Biodiversity Genomics, University of Guelph, 50 Stone Road E, N1G2W1,
9 Guelph, Ontario, Canada

10 ³ Research Department for Limnology, University of Innsbruck, Mondseestr. 9, 5310
11 Mondsee, Austria

12 ⁴ Sinsoma GmbH, Lannes 6, 6176 Voels, Austria

13

14

15 **Keywords:** environmental DNA, aquarium experiment, respirometry, video-analysis, digital
16 PCR

17

18 **Running title:** eDNA shedding in freshwater fish

19

20 **Corresponding Author:**

21 Bettina Thalinger

22 Bettina.Thalinger@gmail.com

23

24 **Abstract**

25 The quantitative measurement of eDNA from field-collected water samples is gaining
26 importance for the monitoring of fish communities and populations. The interpretation of
27 these signal strengths depends, among other factors, on the amount of target eDNA shed
28 into the water. However, shedding rates are presumably associated with species-specific
29 traits such as physiology and behavior. Although such differences between juvenile and adult
30 fish have been previously detected, the general impact of movement and energy use in a
31 resting state on eDNA release into the surrounding water remains hardly addressed.

32 In an aquarium experiment, we compared eDNA shedding between seven fish
33 species occurring in European freshwaters. The investigated salmonids, cyprinids and
34 sculpin exhibit distinct adaptations to microhabitats, diets, and either solitary or schooling
35 behavior. The fish were housed in aquaria with constant water flow and their activity was
36 measured by snapshots taken every 30 s. Water samples for eDNA analysis were taken
37 every 3 h and energy use was determined in an intermittent flow respirometer. After
38 controlling for the effect of fish mass, our results demonstrate a positive correlation between
39 target eDNA quantities as measured with digital PCR, fish activity and energy use, as well as
40 species-specific differences. For cyprinids, the model based on data from individual fish was
41 only partly transferable to groups, which exhibited lower activity and higher energy use.

42 Our findings highlight the importance of fish physiology and behavior for the
43 comparative interpretation of taxon-specific eDNA quantities. Species traits should therefore
44 be incorporated into eDNA-based monitoring and conservation efforts.

45

46 **Introduction**

47 The sensitivity, non-invasiveness and cost-efficiency of environmental DNA (eDNA) based
48 methods has been proven for diverse habitats and species making them powerful new tools
49 for conservation biology and biodiversity assessments (Barnes and Turner, 2016; Deiner et
50 al., 2017; Huerlimann et al., 2020). Regarding the detection of fish species, eDNA-based
51 monitoring outperforms traditional methods such as electrofishing: for example, for the
52 detection of the endangered European weather loach, *Misgurnus fossilis* (Sigsgaard et al.,
53 2015), the assessment of fish communities in Australian streams (McColl-Gausden et al.,
54 2020), and the distribution of brook trout, *Salvelinus fontinalis* in a US watershed (Evans et
55 al., 2017). The manifold successes of eDNA-based species detection lead to a call for more
56 standardization and better reporting practices (Goldberg et al., 2016; Minamoto et al., 2020;
57 Thalinger et al., 2020a) and to an international effort for implementing the technology into
58 routine species monitoring (Leese et al., 2016; Pilliod et al., 2019). Although reporting the
59 presence/absence of particular species is the starting point of these endeavors, a more
60 quantitative interpretation of field-derived eDNA data is key for the broader applicability of
61 this technology.

62 Different processes influence the distribution of eDNA in space and time and the
63 detection probabilities of species from environmental samples, namely the origin,
64 degradation, suspension, resuspension, and transport of eDNA (Barnes and Turner, 2016;
65 Harrison et al., 2019). The latter processes are directly linked to local hydrology (e.g. flow
66 and substrate type (Shogren et al., 2017; Pont et al., 2018; Thalinger et al., 2020b)) and
67 environmental conditions (e.g. water temperature, pH, UV-radiation (Strickler et al., 2015;
68 Lacoursière-Roussel et al., 2016; Tsuji et al., 2017)). The amount of eDNA in the water
69 column is directly linked to fish biomass and originally, this was confirmed for common carp
70 (*Cyprinus carpio*) in an aquarium trial and experimental ponds (Takahara et al., 2012). In
71 subsequent experiments, this positive relationship was confirmed for a range of freshwater
72 and marine fish species (Evans et al., 2016; Lacoursière-Roussel et al., 2016; Sassoubre et

73 al., 2016; Doi et al., 2017; Horiuchi et al., 2019; Jo et al., 2020). However, these results were
74 primarily obtained for individuals at the same life stage.

75 eDNA is released into the environment in the form of mucus, feces, scales, and
76 gametes (Merkes et al., 2014; Barnes and Turner, 2016; Sassoubre et al., 2016; Bylemans
77 et al., 2017). Under natural conditions, differences in fish physiology, diet and behavior are
78 likely to affect this process and confound the interpretation of eDNA-based results from a
79 water body (Klymus et al., 2015). For perch and eel, Maruyama et al. (2014) and Takeuchi et
80 al. (2019), respectively, found lower eDNA shedding rates for adults in comparison to
81 juveniles, which is likely caused by the ontogenetic decrease in metabolic rates (Winberg,
82 1960). However, these findings could not be confirmed in an experiment with a salmonid
83 species (Mizumoto et al., 2018). In general, the metabolic rate and activity differ between fish
84 species due to distinct physiology and behavior with pelagic species being more active and
85 displaying higher resting metabolic rates than benthic species (Johnston et al., 1988; Killen
86 et al., 2010). These factors are frequently hypothesized as underlying causes for spiking
87 eDNA levels at the beginning of aquarium experiments and mismatching quantitative results
88 in studies comparing eDNA levels between species in the same water body (Takahara et al.,
89 2012; Maruyama et al., 2014; Evans et al., 2016).

90 Here, we investigate the effect of fish activity (i.e. movement), energy use (i.e. oxygen
91 use \times oxycaloric factor), and species identity in an aquarium experiment with seven fish
92 species commonly occurring in European rivers and streams. The examined species
93 comprised four salmonids (*Salmo trutta*, *Salvelinus fontinalis*, *Oncorhynchus mykiss*,
94 *Thymallus thymallus*), two cyprinids (*Phoxinus phoxinus*, *Squalius cephalus*) and one sculpin
95 (*Cottus gobio*). *Salmo trutta* is a rhithral species, territorial in later life stages, and primarily
96 feeds on benthic organisms and insects drift on the surface. *Salvelinus fontinalis* and *O.*
97 *mykiss* were anthropogenically introduced into European freshwaters; they are omnivorous
98 and less territorial than *S. trutta*. If possible, these three species choose areas with reduced
99 current close to the main riverbed as preferential microhabitat. *Thymallus thymallus* is also a
100 rhithral species, but its scales are larger and adults primarily use the main riverbed (Spindler,

101 1997; Freyhof and Kottelat, 2007). *Phoxinus phoxinus* is a schooling, small fish species in
102 the rhithral. It feeds on a mixture of plant debris, algae and small invertebrates. The juveniles
103 prefer vegetation-rich microhabitats without current, while adults switch to gravel substrate
104 with low to intermediate flow. *Squalius cephalus* is eurytopic and can occur in habitats with
105 strong to low current. Its juveniles are schooling and omnivorous with adults predominantly
106 preying on fish. *Cottus gobio* is a rheophilic and benthic species primarily feeding on small
107 bottom invertebrates. It has no swim bladder and mostly resides in interstices between large
108 boulders or on coarse gravel characterized by low current (Spindler, 1997; Freyhof and
109 Kottelat, 2007).

110 We hypothesized that higher activity (i.e. movement) leads to higher eDNA
111 concentrations as there is more shearing between the fish surface and the surrounding
112 water, and higher volumes are pumped through the gills due to the elevated oxygen demand.
113 Independent of activity, fish species with higher energy use in a resting state potentially also
114 emit more eDNA. Additionally, the species-specific composition of the constantly renewed
115 cutaneous mucus layer (Ángeles Esteban, 2012) might lead to differences between
116 individual taxa.

117

118

119

120 **Materials and Methods**

121 *Experimental setup*

122 The aquarium experiment was carried out between 2nd March and 17th July 2017 at the
123 Research Department for Limnology Mondsee of the University of Innsbruck, Austria (further
124 on “ILIM”). The salmonid species were purchased from commercial hatcheries, *P. phoxinus*
125 and *S. cephalus* were caught with permission in Lake Mondsee and *C. gobio* were caught
126 with permission in rivers in Tyrol (Austria). Fish individual sizes were chosen as similar as
127 possible within and between species. Until the start of the experiment, the fish species were
128 kept separately in aquaria fed with lake water at the ILIM.

129 In accordance with the regulations of the Austrian Animal Experiment Act (December
130 28, 2012) (Tierversuchsrechtsänderungsgesetz, part 1, section 1, §1, point 2), and with the
131 Directive 2010/63/EU of the European Parliament and of the Council of the European Union
132 (September 22, 2010) on the protection of animals used for scientific purposes (chapter 1,
133 article 1, point 5a), all fish were reared according to regular agriculture (aquaculture)
134 practice, including provision of appropriate tank size, sufficient rate of waterflow, natural
135 photoperiod, ad libitum food supply, and temperatures within the species' thermal tolerance
136 range. This ensured that no pain, suffering, distress or lasting harm was inflicted on the
137 animals, confirmed by the fact that mortality rates were low and equal between rearing
138 groups. Based on the legislative provisions above, no ethics approval and no IACUC protocol
139 was required for the experiments performed. In particular the respirometry experiments were
140 discussed with the legislative authorities (Austrian Federal Ministry of Education, Science
141 and Research and University of Veterinary Medicine, Vienna) and the conclusion was that
142 the assessment of basic metabolism under these conditions (small fish sizes in relatively
143 large chambers) does not incur pain, suffering or distress to the fish and no formal animal
144 experimentation protocol was required.

145 Five aquaria (60 l) and corresponding plastic lids were used in the experiment, each
146 of which was thoroughly cleaned with sodium hypochlorite (5 %) and then rinsed with tap
147 water (fish-DNA free) prior to each experimental run (i.e. changing the fish under
148 investigation). The flow-through rate (aquaria were tap-water fed) was set to 5.45 l/min to
149 mimic natural conditions and keep eDNA concentrations in the fish tanks constant based on
150 previous test runs (data not shown). Water temperature was kept stable at 15 °C for inflowing
151 water automatically stabilizing temperature in the aquaria at this level. Each tank was further
152 equipped with an air-stone to ensure water mixing. At the start of each experimental run, a
153 water sample (negative control) was taken from one of the aquaria and processed as
154 described below. Then, five fish individuals of each species were selected aiming at similar
155 sizes across all fish species. Each fish was placed individually in an aquarium using DNA-
156 free fishnets (Fig. 1). For *P. phoxinus* and *S. cephalus*, the experiment was carried out twice:

157 once with individual fish, and once with groups of three fish per aquarium. The day before the
158 experiment and for its duration, the respective fish were not fed to avoid contamination by
159 fish feed and minimize effects of defecation. Each run started with one day of familiarization
160 in the aquaria.

161

162 *Water sampling, filtration and pH*

163 All equipment used for this process was cleaned with sodium hypochlorite (5 %) and rinsed
164 with tap water prior to each use; DNA-free gloves were always worn. On the second day, 2 l
165 water samples were taken every three hours from 9 o'clock to midnight at the back end of
166 each aquarium (opposite to the inflow) using flexible tubes and 2 l wide neck bottles (Fig. 1).
167 The water samples were immediately filtered in an adjacent laboratory using glass microfiber
168 filters (1.2 µm pore width, 47 mm diameter, Whatman GF/C) and one negative control
169 consisting of 2 l MilliQ-water was included per sampling event. Thereafter, the filters were
170 individually placed in 2 ml reaction tubes and stored at -28 °C until further processing in a
171 special diagnostic molecular laboratory at the Department of Zoology, University of Innsbruck
172 (Austria). After each sampling, pH was measured in three arbitrarily selected fish tanks using
173 a Hach HQ40 device.

174

175 *Activity measurement*

176 During the familiarization time (day 1) and between water samplings, fish swimming activity
177 was quantified using a custom-made activity monitoring system consisting of one high-
178 definition USB camera (Ziggi HD Plus, IPEVO.COM, Great Britain) per aquarium. The
179 cameras were placed at the front of each tank and the focus was set towards the back end
180 (Fig. 1). To enable recordings during the night, aquaria were lighted throughout the two
181 recording days. Additionally, white polystyrene plates were used to cover the bottom and the
182 sides of each fish tank to exclude influences from neighboring aquaria and standardize
183 reflections. The signals from the cameras were acquired with a frame rate of 2 fpm ("frames
184 per minute") with a macro using the image analysis software FIJI (<http://fiji.sc/>, a distribution

185 of ImageJ) for MacOS (Schindelin et al., 2012; Rueden et al., 2017). For each aquarium, a
186 region of interest (“ROI”) excluding the inflow, air-stone and sidewalls was set manually
187 (Supporting Information (SI) 1). Subsequent frames were arithmetically subtracted and the
188 average gray scale within the region of interest, as a quantification of fish activity, was
189 extracted from the difference-images. The dataset was manually checked to exclude artifacts
190 produced by changes in illumination (light/dark illumination of the fish), water sampling,
191 measurement of abiotic factors, fogged-up aquarium front and few other camera movements
192 sometimes leading to a changing region of interest in the recordings (SI 1).

193

194 *Respirometry*

195 A custom made intermittent-flow respirometer was used (Forstner, 1983; Svendsen et al.,
196 2016) including three measurement chambers placed in a larger tank (Fig. 1). The device
197 was cleaned prior to each fish change using a mixture of 3 % hydrogen peroxide and 3 l of
198 tap water. The volume of each chamber was determined prior to the experiment and oxygen
199 saturation (100 %) and temperature (8 – 9 °C) were kept constant in the tank via an airstone
200 and a laboratory heating/cooling device (Lauda DLK 10 and Alpha 1, Lauda Germany). The
201 three chambers of the respirometer were connected to the respirometers’ water circuit,
202 constantly pumping O₂-saturated water from the large tank through the three chambers. For
203 measurements of oxygen consumption, a chamber was cut off from this circuit and a short,
204 closed loop for this chamber was established. Dissolved oxygen was measured in this
205 chamber every 30 s for a period of 15 min. using a YSI ProODO probe (YSI Inc. Yellow
206 Springs, USA) and logged to a computer before the system switched to the next chamber for
207 a 15 min measuring period. On the third day of an experimental run, three of the five fish
208 were placed individually in the chambers avoiding air bubbles and kept there for 24 h for
209 familiarization. On the fourth day, respirometer measurements were carried out for 24 h.
210 Thereafter, the remaining two fish individuals were placed in two measurement chambers for
211 one day of familiarization followed by one day of measurements (day five and six; Fig. 1).
212 The third chamber was left empty, but measured as well, to evaluate potential

213 microorganism-induced oxygen decrease. After the respirometer measurement day, the
214 mass [g] and total length [mm] of each fish was determined before placing them together in a
215 fish tank. For respirometer measurements of fish groups, the three individuals previously
216 sharing an aquarium, were put together in a respirometer chamber.

217

218 *Filter processing and molecular analysis*

219 After defrosting, each filter was soaked with 200 µl of lysis buffer consisting of TES-buffer
220 (0.1 M TRIS, 10 mM EDTA, 2 % sodium dodecyl sulphate; pH 8) and proteinase K (20 mg/ml)
221 in a ratio of 19:1 and incubated at 56 °C over night in a rocking platform. On the next day,
222 filters were transferred with DNA-free forceps to a perforated inset which was repositioned in
223 the top half of the original 2 ml reaction tube and centrifuged for 10 min at 14 000 rpm.
224 Afterwards, filters were discarded and the lysate at the bottom of the reaction tube (300-
225 800 µl) used for DNA extraction. Insets were cleaned in sodium hypochlorite (2.5 %) for at
226 least 30 min, thoroughly washed with MilliQ-water (10 wash steps) and reused.

227 DNA extraction was carried out with the Biosprint 96 instrument (Qiagen) using the
228 Biosprint 96 DNA blood Kit (Qiagen) and the Biosprint 96 tissue extraction protocol in
229 accordance with the manufacturer's instructions except for using 100 µl of TE-buffer instead
230 of AE-buffer for DNA elution. Extractions were carried out in 96-well plates and four negative
231 controls (containing TES-buffer instead of lysate) were included per plate. To process the
232 whole lysate volume, a custom DNA-uptake program was set up: three uptake plates were
233 used and 300 µl of lysate, 300 µl AL-buffer and 300 µl isopropanol were mixed per well in
234 each plate. Missing lysate volumes (i.e. if only a total of 400 µl were available after
235 centrifugation) were replaced by TES-buffer. Additionally, 30 µl MagAttract was added per
236 well in the first plate. Using custom "binding" steps of the robotic platform, the DNA contained
237 in the first plate was transferred to the second one, next a binding step was carried out in the
238 second plate before transferring and releasing the entire collected DNA into the third plate,
239 which was then used for the Biosprint 96 tissue extraction protocol. After extraction, each
240 eluate was transferred to a 1.5 µl reaction tube for subsequent PCR.

241 All used primers (Table 1) have been previously published after extensive specificity
242 and sensitivity testing (Thalinger et al., 2016, 2020b) and additional specificity tests were
243 carried out on the digital PCR (dPCR) system (see below) confirming the specificity of the
244 molecular assays under the following conditions: each 22 μ l dPCR master mix for droplet
245 generation on the QX200 AutoDG (Biorad) consisted of one-time EvaGreen Supermix
246 (Biorad), 0.25 μ M forward and reverse primer (Table 1) and up to 10.5 μ l DNA extract.
247 Depending on the results of initial tests with capillary electrophoresis PCR (i.e. the Relative
248 Fluorescence Units (RFU) of the resulting band; see SI 2), extracts were diluted with
249 molecular grade water for dPCR as follows: RFU < 0.2: undiluted; $0.2 \leq$ RFU < 1.3: 1:1
250 dilution; $1.3 \leq$ RFU < 2: 1:3 dilution; $2 \leq$ RFU: 1:7 dilution. Optimized thermo-cycling
251 conditions were 5 min at 95°C, 40 cycles of 30 s at 95°C, 1 min at 58°C (*O. mykiss*,
252 *P. phoxinus*, and *S. cephalus*) or 60°C (*C. gobio*, *S. fontinalis*, *S. trutta*, and *T. thymallus*),
253 1 min at 72°C, followed by one step of 5 min at 4°C and 5 min at 90°C. dPCR results were
254 analyzed on the QX200™ Droplet Reader and the corresponding QuantaSoft™ Analysis
255 Pro Software (Version 1.7; Biorad). As target signal amplitude varied with the length of the
256 amplified fragment, amplitude thresholds were set individually per primer pair (Table 1) prior
257 to determining target copy numbers per μ l for each DNA extract. Per primer pair, a positive
258 (DNA extract from target species) and a negative control (molecular grade water) were
259 included in dPCR, all of which resulted positive and negative, respectively.

260

261 *Statistical analysis*

262 All calculations and visualizations were carried out in R Version 4.0.2 (R Core Team, 2020)
263 using the packages “ggplot2” (Wickham, 2016), “ggpubr” (Kassambara, 2019), “nlme”
264 (Pinheiro et al., 2020), “AICcmodavg” (Mazerolle, 2020), “rsq” (Zhang, 2020) and “sjPlot”
265 (Lüdecke, 2020). As pH was not measured in all aquaria after each water sampling, missing
266 values were estimated by averaging measurements taken at the respective fish tank before
267 and after the skipped time step. If measurements at the first or last water sampling were
268 missing, the values of the following or previous time step, respectively, were carried over.

269 The cleared activity dataset was visually inspected and summarized for each time
270 step: for example, data obtained during the preceding day were associated with the first
271 eDNA sampling event at 9 AM and measurements between 9 AM and 12 noon were
272 considered relevant for the second water sampling at 12 noon. Mean activity was calculated
273 per time interval. No cleared activity data was available for one *S. trutta* and *S. fontinalis*
274 individual, respectively, and for one *P. phoxinus* and *T. thymallus* individual at a single time
275 step each.

276 The total respirometry dataset was cleared of all 15 min measurement series showing
277 an increase in dissolved O₂. As this value is expected to decrease linearly over the course of
278 a measurement, a linear regression for the oxygen decrease in a measurement chamber
279 over time was calculated for each measurement series. All intervals for which the obtained
280 values showed a bad fit to a linear decrease ($R^2 < 0.8$) were also excluded from further
281 analyses. For each of the remaining measurement intervals, oxygen consumption (OC) in
282 mg / h was calculated as $OC = -s \times 60 \times vol$ where “s” denotes the slope of the linear
283 regression and “vol” the volume of the respective measurement chamber minus the mass of
284 the fish. Per fish species, the obtained value was corrected for the mean oxygen
285 consumption in the empty chamber before calculating total energy use (oxygen consumption
286 $\times 13.6$ J/mg (oxycaloric factor (Brett and Groves, 1979)) per fish. Finally, energy use [J/h]
287 was averaged across the values obtained from individual measurement intervals for each
288 fish and fish group. Due to data clearing, this was not possible for one individual and one
289 group of *C. gobio* and *S. cephalus*, two individuals of *S. fontinalis* and *S. trutta* and three
290 individuals of *T. thymallus*. For these fish, energy use was estimated as the mean of the
291 available values.

292 Concerning the fish-eDNA copy numbers obtained from dPCR, 21 filtered water
293 samples did not lead to an amplification. They were removed from the dataset, as other fish
294 individuals of comparable size and other samplings reliably produced positive results and
295 hence, errors in sample processing might have occurred. One group of *P. phoxinus* had to
296 be excluded from further analyses, as two of three individuals were accidentally chosen

297 *S. cephalus*. To determine whether the pH measurements, mean activity and eDNA copy
298 numbers were significantly influenced by sampling (i.e. time of the day), a one-way repeated
299 measurements ANOVA with rank transformation was calculated for each variable using a
300 combination of fish species and aquarium as random factor. A significant trend could not be
301 detected (Table 2). Despite efforts to standardize the mass of the chosen fish individuals
302 within and between species, fish mass was identified as confounding variable (SI 3). Hence,
303 eDNA copies, mean activity, and energy use were all normalized by the mass of the
304 respective fish individual prior to all further analyses.

305 Generalized Linear Models (GLM) for a Gamma-distributed dependent variable (i.e.
306 eDNA copies; positive continuous variable) and a log-link function were set up to investigate
307 the effects of fish species, energy use, mean activity, and pH (Faraway, 2016). Data
308 obtained from fish groups were excluded from the comparison of model performance. Due to
309 the small number of tested fish individuals and species examined, no models with random
310 factors were considered, but sampling and fish individuals were included in several models to
311 show their ineffectiveness in explaining the obtained data (Table 3). The variable “fish
312 species” was entered via dummy coding into the models using “*C. gobio*” as base category.
313 Corresponding with the focus of this study to investigate the effect of species identity, fish
314 physiology and activity on eDNA shedding, a set of six candidate models was chosen (Table
315 3). AICc, Δ AICc, and AICc weights (ω) were used to evaluate the strength of the six models
316 for describing the data including Nagelkerke, Cragg & Uhler's pseudo-R² values (Burnham
317 and Anderson, 2002). Simulated, scaled residuals were calculated based on the best-
318 performing candidate model, (package: DHARMA (Florian Hartig, 2020); function:
319 “simulateResiduals”; n = 1000); the best performing model passed the consecutive check for
320 outliers and overdispersion.

321 To test the differences between single and grouped fish in the different stages of the
322 experiment, a data subset containing only values obtained from single and grouped
323 *P. phoxinus* and *S. cephalus* was analyzed. Target eDNA copies, energy use, and mean
324 activity (all normalized by fish mass) of the four distinct fish categories were tested for

325 normality and homogeneity of variance with Shapiro-Wilk and Bartlett tests. Then,
326 differences between groups were examined via Kruskal-Wallis tests followed by Wilcoxon
327 rank sum tests with Benjamini-Hochberg-corrected p-values. In a final step, target eDNA
328 copies for groups of *P. phoxinus* and *S. cephalus* were predicted using the model previously
329 established for single fish. Pairwise Wilcoxon tests were used to verify whether there was a
330 significant difference between predicted and measured target eDNA copy numbers for both
331 species separately and combined.

332

333

334

335 **Results**

336 The mean mass of individually housed fish was $3.06 \text{ g} \pm 1.56 \text{ g}$ (SD) and *C. gobio* individuals
337 had the highest mass ($5 \text{ g} \pm 2.1 \text{ g}$ (SD)). Water samples from *P. phoxinus* and *T. thymallus*
338 aquaria had the highest eDNA copy numbers per μl extract and gram fish mass ($31.13 \pm$
339 53.23 (SD) and 47.68 ± 41.13 (SD), respectively; Fig. 2). The normalized mean activity was
340 highest for *S. fontinalis* (1.08 ± 0.33 (SD)) and lowest for *C. gobio* (0.34 ± 0.10 (SD); Fig. 2).
341 The energy use per gram fish mass was highest for *O. mykiss* ($1.81 \text{ J/h} \pm 0.91 \text{ J/h}$ (SD)),
342 while *S. fontinalis* and *S. trutta* aquaria had the lowest pH.

343 The ΔAICc -based comparison of model weight (single fish only) resulted in model #3
344 outperforming five other candidate models (Table 3 and Table 4). Therein, mean activity,
345 energy use and fish species were contained as explanatory variables (pseudo- $R^2_{\text{NCU}} = 0.60$;
346 dispersion parameter = 1.02). Increased activity had a significantly positive effect on eDNA
347 copy numbers ($p < 0.01$) and *P. phoxinus*, *S. cephalus* and *T. thymallus* displayed
348 significantly higher copy numbers compared to *C. gobio* (base group), after controlling for the
349 effect of fish mass. The relationship between energy use and copy numbers was also found
350 to be positive, but not significant ($p = 0.05$; Table 5 and Fig. 3).

351 For single and grouped individuals of *P. phoxinus* and *S. cephalus*, significant
352 differences were detected between the four groups regarding target eDNA copies

353 (Chi² = 10.59; p < 0.05), mean activity (Chi² = 80.95; p < 0.001) and energy use
354 (Chi² = 36.77; p < 0.001): mean activity was significantly higher when fish were kept solitary
355 compared to having them in groups for both species (p < 0.001). Contrastingly, energy use
356 was significantly higher for grouped individuals in both species (p < 0.01). Target eDNA
357 copies were significantly higher for grouped *P. phoxinus* (42.61 ± 48.04 (SD)) compared to
358 single and grouped *S. cephalus* and characterized by few outliers with particularly high eDNA
359 concentration (Fig. 2 and Fig. 4)

360 To test the suitability of model #3 for describing eDNA shedding also for grouped fish,
361 model #3-predicted eDNA copies were compared to the measured copy numbers in the
362 group treatments. For the two species combined, there was no significant difference between
363 predicted and measured copy numbers (W = 0.98, p = 0.34). For *P. phoxinus* alone, no such
364 difference was detected either (W = 270; p = 0.72; Fig. 5), while predicted and measured
365 copy numbers of *S. cephalus* showed a significant difference (W = 614; p < 0.05; Fig. 5).

366

367

368

369 Discussion

370 This experiment confirms the hypothesized positive relationship between eDNA shedding,
371 fish activity and energy use. Additionally, species identity and thereby associated
372 physiological differences were found to influence the amount of released eDNA.
373 Furthermore, our data show that models of eDNA shedding cannot always be generalized
374 from experiments with individual fish to fish groups. For a conclusive habitat-scale estimation
375 of fish communities with eDNA-based methods it is therefore necessary to incorporate
376 species physiology and behavior into the analysis.

377 In early aquarium experiments, the strongest eDNA signals were found right after the
378 introduction of fish into tanks without water circulation and often explained by elevated stress
379 levels through handling and adaption to the new environment (Takahara et al., 2012; Klymus
380 et al., 2015; Maruyama et al., 2019). Hence, many recent studies allow for one or several

381 days of accommodation prior to eDNA sampling (Lacoursière-Roussel et al., 2016; Jo et al.,
382 2019; Takeuchi et al., 2019). We can confirm the positive relationship between fish activity
383 (i.e. movement) and eDNA shedding independent of the introductory phase of an
384 experiment. However, it was not possible to determine the actual reason for the elevated
385 eDNA levels associated with higher activity, as both higher metabolic rates during movement
386 and higher water volumes shearing against the fish body could be responsible for this effect.
387 For eDNA-based field studies this result indicates that signals emitted by highly active fish
388 (e.g. during spawning or predatory behavior) potentially mimic higher levels of fish biomass.

389 Energy use in a resting state as measured with an intermittent-flow respirometer, was
390 also positively correlated with eDNA shedding, albeit with a smaller effect size. Potentially,
391 this finding can be attributed to the higher metabolic rate and larger gill size of active species
392 in combination with higher water volumes pumped through them (Wegner et al., 2009).
393 However, the elevated eDNA signals could also stem from other physiological processes
394 (e.g. defecation), which are known to positively influence eDNA shedding rates (Klymus et
395 al., 2015). As fish were not fed during the entire experiment, the latter factor is potentially
396 negligible. Except for *T. thymallus* the energy use of the species preferring microhabitats with
397 strong currents and preying on fish as adults (primarily *O. mykiss* and *S. cephalus*) was
398 higher than for *C. gobio* and *P. phoxinus*. This is in concordance with general differences in
399 resting metabolic rates between these ecological guilds (Roberts, 1975; Johnston et al.,
400 1988; Killen et al., 2010).

401 There were distinct differences in eDNA shedding between the species, with
402 *T. thymallus*, *P. phoxinus* and *S. cephalus* emitting the most eDNA. The adaptation to
403 habitats with stronger currents (Freyhof and Kottelat, 2007), namely an increased mucus
404 production in combination with comparably large scales, might explain this result for
405 *T. thymallus* and *S. cephalus*. The underlying taxonomy could also contribute to this pattern if
406 cyprinids (*P. phoxinus* and *S. cephalus*) generally pass more DNA into the surrounding water
407 via their gills, feces or mucus. Another explanation for the high eDNA shedding of cyprinids
408 in this experiment could be the stress induced by solitary housing. The model estimating

409 eDNA concentrations for individual fish could not fully explain the findings obtained for
410 grouped fish: the activity of both *P. phoxinus* and *S. cephalus* was significantly lower when
411 fish were held in groups, while their energy use was significantly higher. However, the
412 opposing direction of these two effects potentially explains why differences between
413 predicted and measured eDNA quantities were only significant for *S. cephalus*.

414 Generally, the measured eDNA concentrations per μl DNA extract were right-skewed
415 and a few exceptionally high values showed considerable influence on the size of standard
416 deviations. These results were independent of fish handling and stress during the
417 introduction phase as eDNA sampling started only after 24 h and the aquaria had constant
418 flow with the entire volume being renewed every 11 min. Such “outliers” were also detected
419 in other aquarium experiments (Klymus et al., 2015; Wilcox et al., 2016) and cell-
420 conglomerates released into the surrounding water were previously deemed responsible for
421 this pattern (Wilcox et al., 2016). Additionally, the size distribution of eDNA particles (from
422 $< 0.2 \mu\text{m}$ to $>180 \mu\text{m}$) and commonly detected fragment sizes suggest intact cells or
423 organelles as the primary source of eDNA in the water column (reviewed by Harrison et al.,
424 2019). Our data support the hypothesis of constant eDNA shedding rates at constant
425 environmental conditions. We could not observe any effects of sampling time, possibly due to
426 the constant illumination of the aquaria. Hence, this aspect is not necessarily transferable to
427 natural environments where fish are known to exhibit distinct diurnal movement patterns
428 (Helfman, 1986).

429 The influence of fish mass on eDNA concentrations was not in the focus of this
430 experiment and fish individuals were as similar in size/mass as possible. However, adult fish
431 of *P. phoxinus* and *C. gobio* are considerably smaller in comparison to the other species
432 (Freyhof and Kottelat, 2007) and the respective juveniles were thus in a later life stage.
433 Based on the ontogenetic decrease in metabolic rate (Winberg, 1960), this could be an
434 alternative explanation for the low energy use of these two species. For studies investigating
435 eDNA shedding directly from live animals, biomass will always be an influential and
436 potentially confounding variable and should thus be considered carefully already during

437 experimental design. Depending on the actual main source of eDNA (feces, mucus...) it is
438 furthermore questionable, if fish mass is the best index variable to describe study animals.
439 Fish length should also be considered as it is a good approximator for activity measurements
440 via videotaping, an estimator of fish surface, and at least for eels a good estimator of eDNA
441 shedding (Takeuchi et al., 2019). Since body shapes and fins differ a lot between taxa
442 (Freyhof and Kottelat, 2007), it could also be beneficial to factor body shape (fusiform,
443 elongated, flat,...) into future experiments in case the fish surface is the primary source of
444 eDNA. Finally, individual differences are well documented for fish behavior and metabolic
445 rates (Metcalf et al., 2016). The number of study animals in future experiments should thus
446 be increased to better control for such effects within the same species.

447 Our results demonstrate that for the successful application of eDNA-based methods
448 on a habitat scale it is necessary to incorporate fish physiology and behavior not only in the
449 study design and sampling process (e.g. by sampling at different depths and in different
450 micro-habitats (Littlefair et al., 2020)), but also during data analysis (Barnes and Turner,
451 2016; Thalinger et al., 2020a). Seasonal patterns could have a much stronger effect on
452 eDNA concentrations in the water column as previously assumed: for instance, many
453 cyprinids in European freshwaters seek calm areas without current during the winter. Their
454 eDNA is less likely to spread through the water column and additionally, their decreased
455 activity and energy use lower the detection probability even further. In the future, the eDNA
456 shedding of manifold fish species and families in relation to their biomass, activity and energy
457 use should be investigated to deepen our understanding of taxon-specific effects. Until then,
458 estimations of fish biomass from eDNA quantities in field-collected samples should at least
459 take distinct physiology and behavior into account, especially for comparative analyses
460 between species or seasons.

461

462

463 **Acknowledgements**

464 This research was conducted within the eDNA-Alpfish project funded by the Austria
465 Research Promotion Agency (FFG); project number 853219. We thank R. Vogt for his
466 support during the experiment, M. Böcker for assistance with the background literature, J.
467 Harvie for input on the statistic analysis, and C. Moritz and D. Kirschner for their help in
468 obtaining the *C. gobio* individuals.

469

470

471 **Conflict of interest**

472 MT is the co-founder of Sinsoma GmbH, a for profit company dedicated to DNA analyses in
473 environmental studies.

474

475

476 **Author contribution statement**

477 MT and JW, conceived the study; the experiment was designed by BT, MT, TS and JW, and
478 carried out by AR and AT under the supervision of BT and JW. TS was responsible for the
479 processing of activity data; AR, AT and YP were responsible for laboratory processing of the
480 eDNA samples under the supervision of BT who also carried out statistical analysis and
481 wrote the first draft of the manuscript which was revised by all co-authors.

482

483

484 **Data Availability Statement**

485 All data on eDNA signals, fish activity, energy use, fish mass and pH have been uploaded to
486 Figshare and are available at <https://doi.org/10.6084/m9.figshare.13151180.v1>

487

488 **References**

- 489 Ángeles Esteban, M. (2012). An Overview of the Immunological Defenses in Fish Skin. *ISRN*
490 *Immunol.* 2012, 1–29. doi:10.5402/2012/853470.
- 491 Barnes, M. A., and Turner, C. R. (2016). The ecology of environmental DNA and implications
492 for conservation genetics. *Conserv. Genet.* 17, 1–17. doi:10.1007/s10592-015-0775-4.
- 493 Brett, J. R., and Groves, T. D. D. (1979). “Physiological energetics,” in *Fish Physiology Vol.*
494 *8, Bioenergetics and Growth.*, eds. W. S. Hoar, D. J. Randall, and J. R. Brett (New York:
495 Academic Press). Available at: <https://trove.nla.gov.au/work/18117950>.
- 496 Burnham, K. P., and Anderson, D. R. (2002). *Model Selection and Multimodel Inference.* 2nd
497 ed. New York: Springer doi:10.1007/b97636.
- 498 Bylemans, J., Furlan, E. M., Hardy, C. M., McGuffie, P., Lintermans, M., and Gleeson, D. M.
499 (2017). An environmental DNA-based method for monitoring spawning activity: a case
500 study, using the endangered Macquarie perch (*Macquaria australasica*). *Methods Ecol.*
501 *Evol.* 8, 646–655. doi:10.1111/2041-210X.12709.
- 502 Deiner, K., Bik, H. M., Mächler, E., Seymour, M., Lacoursière-Roussel, A., Altermatt, F., et al.
503 (2017). Environmental DNA metabarcoding: Transforming how we survey animal and
504 plant communities. *Mol. Ecol.* 26, 5872–5895. doi:10.1111/mec.14350.
- 505 Doi, H., Inui, R., Akamatsu, Y., Kanno, K., Yamanaka, H., Takahara, T., et al. (2017).
506 Environmental DNA analysis for estimating the abundance and biomass of stream fish.
507 *Freshw. Biol.* 62, 30–39. doi:10.1111/fwb.12846.
- 508 Evans, N. T., Olds, B. P., Renshaw, M. A., Turner, C. R., Li, Y., Jerde, C. L., et al. (2016).
509 Quantification of mesocosm fish and amphibian species diversity via environmental
510 DNA metabarcoding. *Mol. Ecol. Resour.* 16, 29–41. doi:10.1111/1755-0998.12433.
- 511 Evans, N. T., Shirey, P. D., Wieringa, J. G., Mahon, A. R., and Lamberti, G. A. (2017).
512 Comparative Cost and Effort of Fish Distribution Detection via Environmental DNA
513 Analysis and Electrofishing. *Fisheries* 42, 90–99. doi:10.1080/03632415.2017.1276329.
- 514 Faraway, J. J. (2016). *Extending the Linear Model with R. Generalized Linear, Mixed Effects*
515 *and Nonparametric Regression Models.* 2nd ed. New York: Chapman and Hall/CRC

- 516 doi:10.1201/9781315382722.
- 517 Florian Hartig (2020). DHARMa: Residual Diagnostics for Hierarchical (Multi-Level / Mixed)
518 Regression ModelsNo Title. Available at: <https://cran.r-project.org/package=DHARMa>.
- 519 Forstner, H. (1983). "An Automated Multiple-Chamber Intermittent-Flow Respirometer," in
520 *Polarographic Oxygen Sensors* (Springer Berlin Heidelberg), 111–126. doi:10.1007/978-
521 3-642-81863-9_12.
- 522 Freyhof, J., and Kottelat, M. (2007). *Handbook of European freshwater fishes* . Springer
523 Science and Business Media LLC doi:10.1007/s10228-007-0012-3.
- 524 Goldberg, C. S., Turner, C. R., Deiner, K., Klymus, K. E., Thomsen, P. F., Murphy, M. A., et
525 al. (2016). Critical considerations for the application of environmental DNA methods to
526 detect aquatic species. *Methods Ecol. Evol.* 7, 1299–1307. doi:10.1111/2041-
527 210X.12595.
- 528 Harrison, J. B., Sunday, J. M., and Rogers, S. M. (2019). Predicting the fate of eDNA in the
529 environment and implications for studying biodiversity. *Proc. R. Soc. B Biol. Sci.* 286.
530 doi:10.1098/rspb.2019.1409.
- 531 Helfman, G. S. (1986). "Fish Behaviour by Day, Night and Twilight," in *The Behaviour of*
532 *Teleost Fishes* (Springer US), 366–387. doi:10.1007/978-1-4684-8261-4_14.
- 533 Horiuchi, T., Masuda, R., Murakami, H., Yamamoto, S., and Minamoto, T. (2019).
534 Biomass-dependent emission of environmental DNA in jack mackerel *Trachurus*
535 *japonicus* juveniles. *J. Fish Biol.* 95, jfb.14095. doi:10.1111/jfb.14095.
- 536 Huerlimann, R., Cooper, M. K., Edmunds, R. C., Villacorta-Rath, C., Le Port, A., Robson, H.
537 L. A., et al. (2020). Enhancing tropical conservation and ecology research with aquatic
538 environmental DNA methods: an introduction for non-environmental DNA specialists.
539 *Anim. Conserv.* doi:10.1111/acv.12583.
- 540 Jo, T., Arimoto, M., Murakami, H., Masuda, R., and Minamoto, T. (2020). Estimating
541 shedding and decay rates of environmental nuclear DNA with relation to water
542 temperature and biomass. *Environ. DNA* 2, 140–151. doi:10.1002/edn3.51.
- 543 Jo, T., Murakami, H., Yamamoto, S., Masuda, R., and Minamoto, T. (2019). Effect of water

- 544 temperature and fish biomass on environmental DNA shedding, degradation, and size
545 distribution. *Ecol. Evol.* 9, 1135–1146. doi:10.1002/ece3.4802.
- 546 Johnston, I. A., Camm, J. P., and White, M. (1988). Specialisations of swimming muscles in
547 the pelagic antarctic fish *Pleuragramma antarcticum*. *Mar. Biol.* 100, 3–12.
548 doi:10.1007/BF00392949.
- 549 Kassambara, A. (2019). ggpubr: “ggplot2” Based Publication Ready Plots. Available at:
550 <https://cran.r-project.org/package=ggpubr>.
- 551 Killen, S. S., Atkinson, D., and Glazier, D. S. (2010). The intraspecific scaling of metabolic
552 rate with body mass in fishes depends on lifestyle and temperature. *Ecol. Lett.* 13, 184–
553 193. doi:10.1111/j.1461-0248.2009.01415.x.
- 554 Klymus, K. E., Richter, C. A., Chapman, D. C., and Paukert, C. (2015). Quantification of
555 eDNA shedding rates from invasive bighead carp *Hypophthalmichthys nobilis* and silver
556 carp *Hypophthalmichthys molitrix*. *Biol. Conserv.* 183, 77–84.
557 doi:10.1016/j.biocon.2014.11.020.
- 558 Lacoursière-Roussel, A., Rosabal, M., and Bernatchez, L. (2016). Estimating fish abundance
559 and biomass from eDNA concentrations: variability among capture methods and
560 environmental conditions. *Mol. Ecol. Resour.* 16, 1401–1414. doi:10.1111/1755-
561 0998.12522.
- 562 Leese, F., Altermatt, F., Bouchez, A., Ekrem, T., Hering, D., Meissner, K., et al. (2016).
563 DNAqua-Net: Developing new genetic tools for bioassessment and monitoring of
564 aquatic ecosystems in Europe. *Res. Ideas Outcomes* 2, e11321.
565 doi:10.3897/rio.2.e11321.
- 566 Littlefair, J. E., Hrenchuk, L. E., Blanchfield, P. J., Rennie, M. D., and Cristescu, M. E. (2020).
567 Thermal stratification and fish thermal preference explain vertical eDNA distributions in
568 lakes. *Mol. Ecol.*, mec.15623. doi:10.1111/mec.15623.
- 569 Lüdecke, D. (2020). sjPlot: Data Visualization for Statistics in Social Science. Available at:
570 <https://cran.r-project.org/package=sjPlot>.
- 571 Maruyama, A., Nakamura, K., Yamanaka, H., Kondoh, M., and Minamoto, T. (2014). The

- 572 Release Rate of Environmental DNA from Juvenile and Adult Fish. *PLoS One* 9,
573 e114639. doi:10.1371/journal.pone.0114639.t001.
- 574 Maruyama, A., Nakamura, K., Yamanaka, H., Kondoh, M., and Minamoto, T. (2019).
575 Correction: The release rate of environmental DNA from juvenile and adult fish (PLoS
576 ONE (2019) 14:2 (e0212145) DOI: 10.1371/journal.pone.0114639). *PLoS One* 14.
577 doi:10.1371/journal.pone.0212145.
- 578 Mazerolle, M. J. (2020). AICcmodavg: Model selection and multimodel inference based on
579 (Q)AIC(c). Available at: <https://cran.r-project.org/package=AICcmodavg>.
- 580 McColl-Gausden, E., Weeks, A., Coleman, R., Robinson, K., Song, S., Raadik, T., et al.
581 (2020). Multi-species models reveal that eDNA metabarcoding is more sensitive than
582 backpack electrofishing for conducting fish surveys in freshwater streams. *Mol. Ecol.*
583 doi:10.1111/mec.15644.
- 584 Merkes, C. M., McCalla, S. G., Jensen, N. R., Gaikowski, M. P., and Amberg, J. J. (2014).
585 Persistence of DNA in Carcasses, Slime and Avian Feces May Affect Interpretation of
586 Environmental DNA Data. *PLoS One* 9, e113346. doi:10.1371/journal.pone.0113346.
- 587 Metcalfe, N. B., Van Leeuwen, T. E., and Killen, S. S. (2016). Does individual variation in
588 metabolic phenotype predict fish behaviour and performance? *J. Fish Biol.* 88, 298–321.
589 doi:10.1111/jfb.12699.
- 590 Minamoto, T., Miya, M., Sado, T., Seino, S., Doi, H., Kondoh, M., et al. (2020). An illustrated
591 manual for environmental DNA research: Water sampling guidelines and experimental
592 protocols. *Environ. DNA*, edn3.121. doi:10.1002/edn3.121.
- 593 Mizumoto, H., Urabe, H., Kanbe, T., Fukushima, M., and Araki, H. (2018). Establishing an
594 environmental DNA method to detect and estimate the biomass of Sakhalin taimen, a
595 critically endangered Asian salmonid. *Limnology* 19, 219–227. doi:10.1007/s10201-017-
596 0535-x.
- 597 Pilliod, D. S., Laramie, M. B., MacCoy, D., and Maclean, S. (2019). Integration of
598 eDNA-Based Biological Monitoring within the U.S. Geological Survey's National
599 Streamgauge Network. *JAWRA J. Am. Water Resour. Assoc.* 55, 1505–1518.

- 600 doi:10.1111/1752-1688.12800.
- 601 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., and R Core Team (2020). nlme: Linear and
602 Nonlinear Mixed Effects Models. Available at: <https://cran.r-project.org/package=nlme>.
- 603 Pont, D., Rocle, M., Valentini, A., Civade, R., Jean, P., Maire, A., et al. (2018). Environmental
604 DNA reveals quantitative patterns of fish biodiversity in large rivers despite its
605 downstream transportation. *Sci. Rep.* 8. doi:10.1038/s41598-018-28424-8.
- 606 R Core Team (2020). R: A Language and Environment for Statistical Computing. Available
607 at: <https://www.r-project.org/>.
- 608 Roberts, J. L. (1975). Active branchial and ram gill ventilation in fishes. *BIOL.BULL.* 148, 85–
609 105. doi:10.2307/1540652.
- 610 Rueden, C. T., Schindelin, J., Hiner, M. C., DeZonia, B. E., Walter, A. E., Arena, E. T., et al.
611 (2017). ImageJ2: ImageJ for the next generation of scientific image data. *BMC*
612 *Bioinformatics* 18, 1–26. doi:10.1186/s12859-017-1934-z.
- 613 Sassoubre, L. M., Yamahara, K. M., Gardner, L. D., Block, B. A., and Boehm, A. B. (2016).
614 Quantification of Environmental DNA (eDNA) Shedding and Decay Rates for Three
615 Marine Fish. *Environ. Sci. Technol.* 50, 10456–10464. doi:10.1021/acs.est.6b03114.
- 616 Schindelin, J., Arganda-Carreras, I., Frise, E., Kaynig, V., Longair, M., Pietzsch, T., et al.
617 (2012). Fiji: An open-source platform for biological-image analysis. *Nat. Methods* 9,
618 676–682. doi:10.1038/nmeth.2019.
- 619 Shogren, A. J., Tank, J. L., Andruszkiewicz, E., Olds, B., Mahon, A. R., Jerde, C. L., et al.
620 (2017). Controls on eDNA movement in streams: Transport, Retention, and
621 Resuspension. *Sci. Rep.* 7, 1–11. doi:10.1038/s41598-017-05223-1.
- 622 Sigsgaard, E. E., Carl, H., Møller, P. R., and Thomsen, P. F. (2015). Monitoring the near-
623 extinct European weather loach in Denmark based on environmental DNA from water
624 samples. *Biol. Conserv.* 183, 46–52. doi:10.1016/j.biocon.2014.11.023.
- 625 Spindler, T. (1997). *Fischfauna in Österreich. Ökologie – Gefährdung – Bioindikation –*
626 *Fischerei – Gesetzgebung*. Umweltbund. Vienna, Austria: Umweltbundesamt.
- 627 Strickler, K. M., Fremier, A. K., and Goldberg, C. S. (2015). Quantifying effects of UV-B,

- 628 temperature, and pH on eDNA degradation in aquatic microcosms. *Biol. Conserv.* 183,
629 85–92. doi:10.1016/j.biocon.2014.11.038.
- 630 Svendsen, M. B. S., Bushnell, P. G., and Steffensen, J. F. (2016). Design and setup of
631 intermittent-flow respirometry system for aquatic organisms. *J. Fish Biol.* 88, 26–50.
632 doi:10.1111/jfb.12797.
- 633 Takahara, T., Minamoto, T., Yamanaka, H., Doi, H., and Kawabata, Z. (2012). Estimation of
634 Fish Biomass Using Environmental DNA. *PLoS One* 7, e35868.
635 doi:10.1371/journal.pone.0035868.
- 636 Takeuchi, A., Iijima, T., Kakuzen, W., Watanabe, S., Yamada, Y., Okamura, A., et al. (2019).
637 Release of eDNA by different life history stages and during spawning activities of
638 laboratory-reared Japanese eels for interpretation of oceanic survey data. *Sci. Rep.* 9,
639 1–9. doi:10.1038/s41598-019-42641-9.
- 640 Thalinger, B., Deiner, K., Harper, L., Rees, H., Blackman, R., Sint, D., et al. (2020a). A
641 validation scale to determine the readiness of environmental DNA assays for routine
642 species monitoring. *bioRxiv*, 2020.04.27.063990. doi:10.1101/2020.04.27.063990.
- 643 Thalinger, B., Kirschner, D., Pütz, Y., Moritz, C., Schwarzenberger, R., Wanzenböck, J., et
644 al. (2020b). Lateral and longitudinal fish eDNA distribution in dynamic riverine habitats.
645 *bioRxiv*, 2020.05.28.120147. doi:10.1101/2020.05.28.120147.
- 646 Thalinger, B., Oehm, J., Mayr, H., Obwexer, A., Zeisler, C., and Traugott, M. (2016).
647 Molecular prey identification in Central European piscivores. *Mol. Ecol. Resour.* 16,
648 123–137. doi:10.1111/1755-0998.12436.
- 649 Tsuji, S., Ushio, M., Sakurai, S., Minamoto, T., and Yamanaka, H. (2017). Water
650 temperature-dependent degradation of environmental DNA and its relation to bacterial
651 abundance. *PLoS One* 12. doi:10.1371/journal.pone.0176608.
- 652 Wegner, N. C., Sepulveda, C. A., Bull, K. B., and Graham, J. B. (2009). Gill morphometrics in
653 relation to gas transfer and ram ventilation in high-energy demand teleosts: Scombrids
654 and billfishes. *J. Morphol.* 271, 36–49. doi:10.1002/jmor.10777.
- 655 Wickham, H. (2016). *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York

656 Available at: <https://ggplot2.tidyverse.org>.

657 Wilcox, T. M., McKelvey, K. S., Young, M. K., Sepulveda, A. J., Shepard, B. B., Jane, S. F.,
658 et al. (2016). Understanding environmental DNA detection probabilities: A case study
659 using a stream-dwelling char *Salvelinus fontinalis*. *Biol. Conserv.* 194, 209–216.
660 doi:10.1016/j.biocon.2015.12.023.

661 Winberg, G. G. (1960). *Rate of metabolism and food requirements offishes.* , eds. F. E. J. Fry
662 and W. E. Ricker Fish.Res. Board Can. Transl. Ser. No. 194.

663 Zhang, D. (2020). rsq: R-Squared and Related Measures. Available at: [https://cran.r-](https://cran.r-project.org/package=rsq)
664 [project.org/package=rsq](https://cran.r-project.org/package=rsq).

665

666

667 Tables and Figures

668

669 **Table 1:** PCR assays used to amplify fish-eDNA: columns denote the target taxon of each
670 primer combination, primer names, sequences, their respective concentration in PCR, target
671 gene, amplicon sizes, and threshold value for positive droplets in dPCR. Additionally, the
672 source column states whether a primer pair has been previously published. Please note that
673 the *Salvelinus* sp. primer pair was designed to amplify both *S. fontinalis* and *Salvelinus*
674 *umbla*.

675

676 **Table 2:** The results of one-way repeated measures ANOVA with rank transformation
677 examining a potential effect of sampling on pH, mean activity and target eDNA copy
678 numbers.

679

680 **Table 3:** Covariate structures of the candidate models compared for their potential to explain
681 the target eDNA copy numbers per gram fish and μl extract obtained from single-fish aquaria
682 with the following parameters: fish species identity (seven species), mean activity (per gram
683 fish), energy use (per gram fish) and pH. Sampling event and fish individual number were
684 included primarily to show their insignificance.

685

686 **Table 4:** Results of the ordinal ranking based on ΔAICc for the Generalized Linear Models
687 (Table 3). Models are ordered from high to low weight and K denotes for the number of
688 estimable parameters, AICc for the second-order variant of Akaike's Information Criterion,
689 ΔAICc for AICc difference and ω for Akaike weight.

690

691 **Table 5:** The highest weight ($\omega = 0.90$) Generalized Linear Model (#3) describing the
692 measured eDNA copy numbers via mean activity, energy use and fish species identity.
693 Significant p-values of fish species in the model refer to a significant difference between
694 *Cottus gobio* (used as base category for dummy coding) and the respective fish species.

695

696 **Figure 1:** The setup of the aquarium experiment carried out with seven fish species: five
697 individual fish were put in fish tanks for water sampling (eDNA) and activity recordings (day 1
698 and 2) followed by respirometer measurements (three individuals on days 3 and 4; two
699 individuals plus empty control chamber on days 5 and 6). For *Phoxinus phoxinus* and
700 *Squalius cephalus* the experiment was repeated using groups of three individuals.

701

702 **Figure 2:** Key parameters obtained during the experiment for single fish. Boxplots display
703 target eDNA copies per μl extract, energy use [J/h], mean activity, and pH per fish species.
704 Fish species are abbreviated: "Cot gob": *Cottus gobio*; "Onc myk": *Oncorhynchus mykiss*;
705 "Pho pho": *Phoxinus phoxinus*; "Sal fon": *Salvelinus fontinalis*; "Sal tru": *Salmo trutta*; "Squ
706 cep": *Squalius cephalus*; "Thy thy": *Thymallus thymallus*. The variables target eDNA copies,
707 mean activity and energy use were normalized by fish mass to control for the effect of this
708 confounding variable.

709

710 **Figure 3:** Graphic representation of the GLM model coefficients (model #3) best describing
711 the obtained target eDNA copy numbers. Significance codes of denoted fish species indicate
712 differences in comparison to the base category *Cottus gobio*, whiskers denote the 95%-CI.

713

714 **Figure 4:** Comparison of target eDNA copies, mean activity and energy use (normalized by
715 fish mass) in aquaria obtained from single and grouped individuals of *Phoxinus phoxinus* and
716 *Squalius cephalus*. Different lower case letters above boxplots code for significant
717 differences ($p < 0.05$) between categories, which are abbreviated as: “Pho pho”: *Phoxinus*
718 *phoxinus* (single fish); “Pho pho g”: *Phoxinus phoxinus* grouped fish; “Squ cep”: *Squalius*
719 *cephalus* (single fish); “Squ cep g”: *Squalius cephalus* grouped fish.

720

721 **Figure 5:** For groups of *Phoxinus phoxinus* (Pho pho g) and *Squalius cephalus* (Squ cep g)
722 measured and predicted copy numbers are plotted: left: against each other; middle: predicted
723 copy numbers are compared between species; right: comparison of measured copy numbers
724 between the two species. The measured copy numbers were log-transformed to enable a
725 direct comparison with the values predicted by the Gamma GLM with log-link function. For
726 *S. cephalus* a significant difference between measured and predicted copy numbers was
727 detected ($t = 2.37$; $p = 0.02$).

728 **Table 1**

729
730
731
732
733
734
735
736
737
738
739
740
741
742
743
744
745
746
747

target taxon	primer name	primer sequence (5' - 3')	primer conc. in PCR (μM)	target gene	fragment length (bp)	amplitude threshold (dPCR)	source
<i>Cottus gobio</i>	Cot-gob-S632	GAATAAAGGACTAAACCAAGTGGG	0.25	16S	118	13,500	Thalinger et al. 2016
	Cot-gob-A641	GCTGTAGCTCTCAGTTGTAGGAAAA	0.25				
<i>Salmo trutta</i>	Sal-tru-S1002	TCTCTTGATTCGGGCAGAACTC	0.25	COI	89	8,400	Thalinger et al. 2020
	Sal-tru-A1002	CGAAGGCATGGGCTGTAACA	0.25				
<i>Oncorhynchus mykiss</i>	Onc-myk-S655	TCTCCCTTCATTTAGCTGGAATC	0.25	COI	82	12,500	Thalinger et al. 2016
	Onc-myk-S655	GCTGGAGGTTTTATGTTAATAATGGTC	0.25				
<i>Salvelinus</i> spp.	Sal-vel-S651	ATAGTCGGCACCGCCCTT	0.25	COI	112	14,000	Thalinger et al. 2016
	Sal-vel-A651	TAACGAAGGCATGGGCTGTT	0.25				
<i>Thymallus thymallus</i>	Thy-thy-S653	ATCAAATTTATAATGTGATCGTCACG	0.25	COI	179	14,000	Thalinger et al. 2016
	Thy-thy-A653	AAGAAAGGACGGGGGAAGC	0.25				
<i>Phoxinus phoxinus</i>	Pho-pho-S639	CGTGCAGAAGCGGATATAAATAC	0.25	16S	128	15,750	Thalinger et al. 2016
	Pho-pho-A648	CCAACCGAAGGTAAGTCTTATTG	0.25				
<i>Squalius cephalus</i>	Squ-cep-S669	CAGTATACCCACCGCTTGCG	0.25	COI	130	14,250	Thalinger et al. 2016
	Squ-cep-A669	TTAATAATTGTGGTAATGAAGTTGACC	0.25				

Table 2

		F-value	p-value
pH	(intercept)	34.52	< 0.001
	sampling	2.22	0.053
mean activity	(intercept)	78.78	<0.001
	sampling	0.76	0.57
target copies per μ l	(intercept)	39.00	<0.001
	sampling	0.38	0.86

Table 3

model #	covariate structures
1	mean activity + energy use + fish species + pH + sampling + fish individual
2	mean activity + energy use + fish species + sampling + fish individual
3	mean activity + energy use + fish species
4	mean activity + energy use
5	mean activity + fish species
6	fish species

Table 4

model #	K	AICc	ΔAICc	ω	pseudo-R^2_{NCU}
3	10	1208.32	0	0.90	0.60
5	9	1212.76	4.44	0.10	0.58
2	19	1221.36	13.04	0	0.62
6	8	1223.20	14.88	0	0.55
1	20	1223.64	15.32	0	0.62
4	4	1346.99	138.67	0	0.04

Table 5

predictor variable	parameter estimate	standard error	lower 95% CI	upper 95% CI	t-value	p-value
intercept	0.26	0.26	-0.23	0.77	0.97	0.33
mean activity	1.05	0.36	0.36	1.75	2.91	< 0.01**
energy use [J/h]	0.40	0.20	0.04	0.77	1.97	0.05
<i>Oncorhynchus mykiss</i>	0.72	0.37	0.06	1.41	1.96	0.05
<i>Phoxinus phoxinus</i>	2.33	0.29	1.76	2.91	7.90	< 0.001***
<i>Salvelinus fontinalis</i>	0.53	0.39	-0.23	1.31	1.34	0.18
<i>Salmo trutta</i>	0.28	0.31	-0.34	0.92	0.91	0.36
<i>Squalius cephalus</i>	1.16	0.32	0.54	1.80	3.65	< 0.001**
<i>Thymallus thymallus</i>	2.96	0.28	2.43	3.49	10.51	< 0.001***

Figure 1

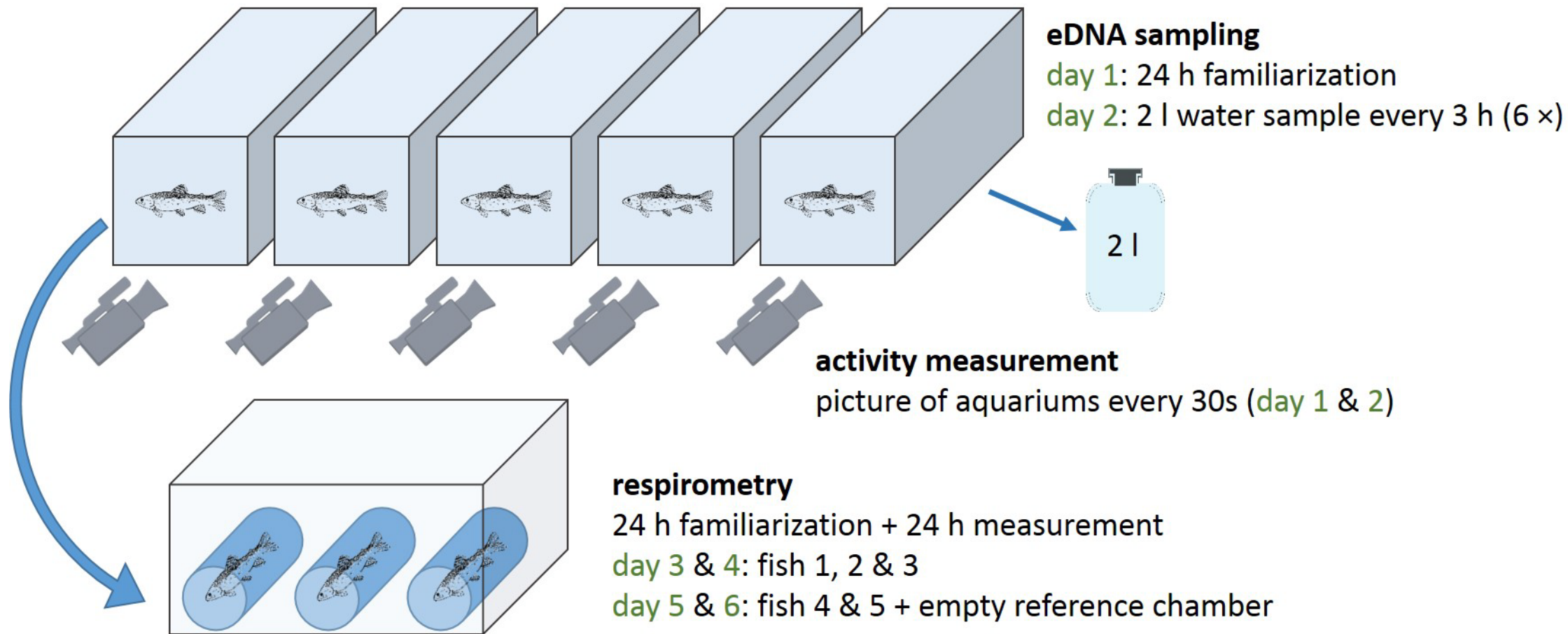


Figure 2

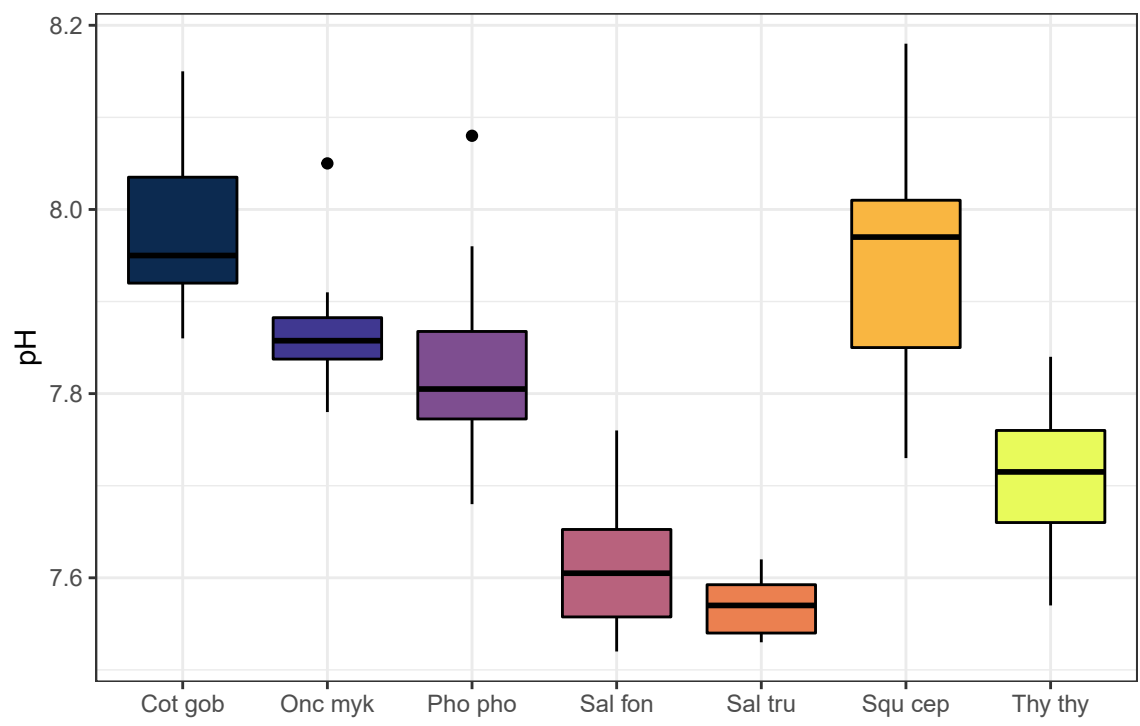
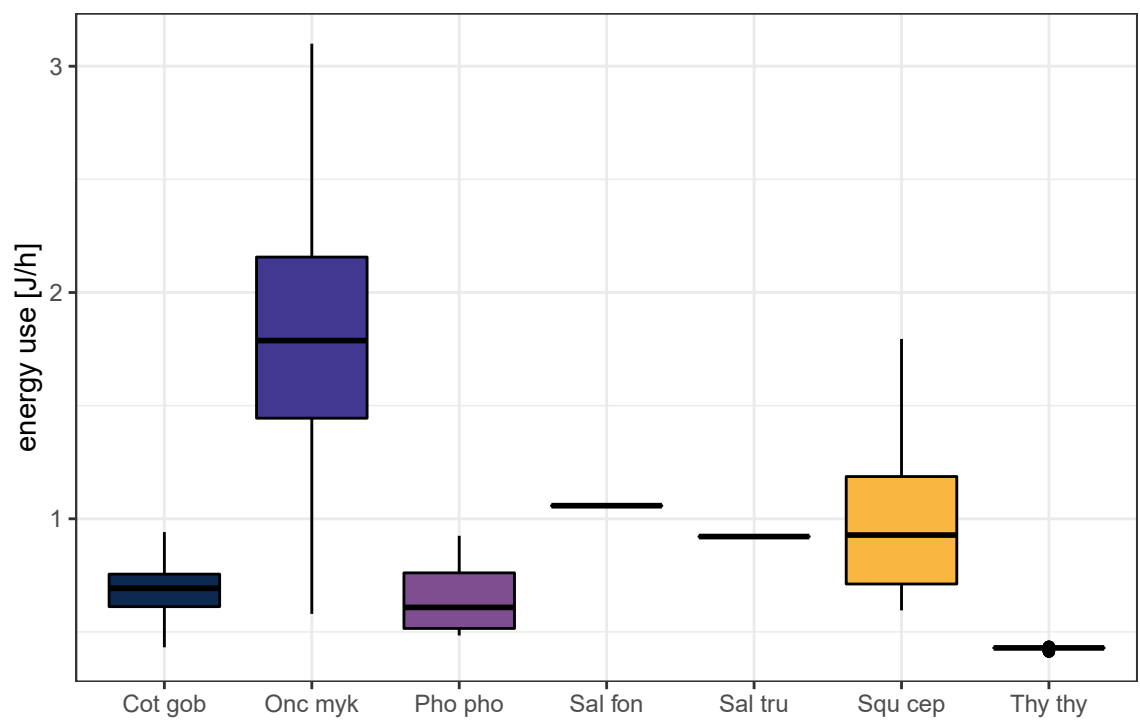
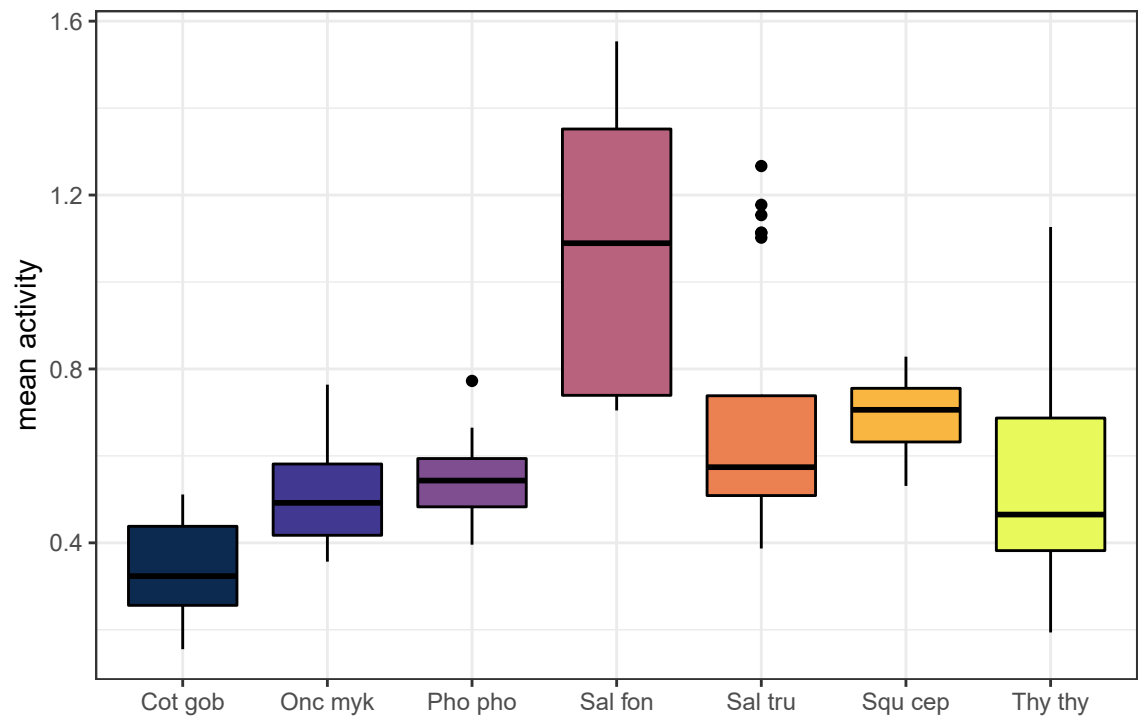
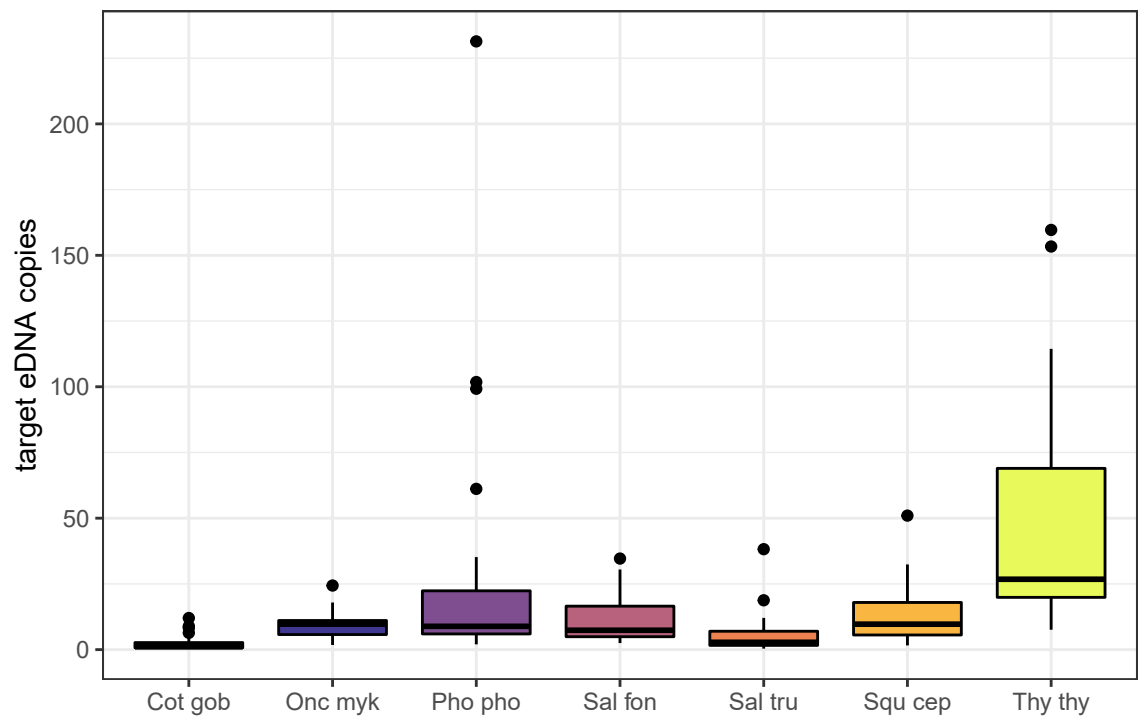


Figure 3

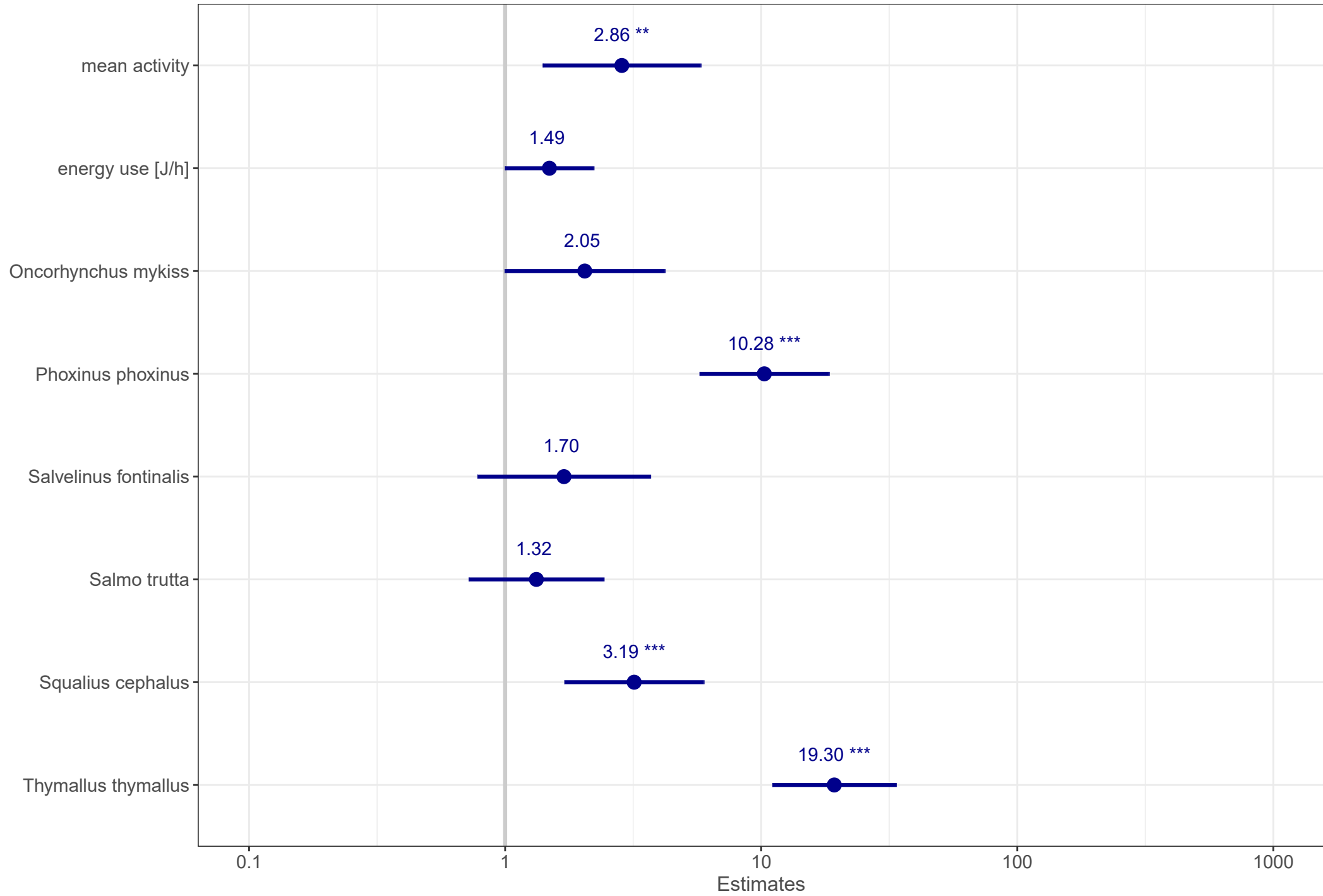


Figure 4

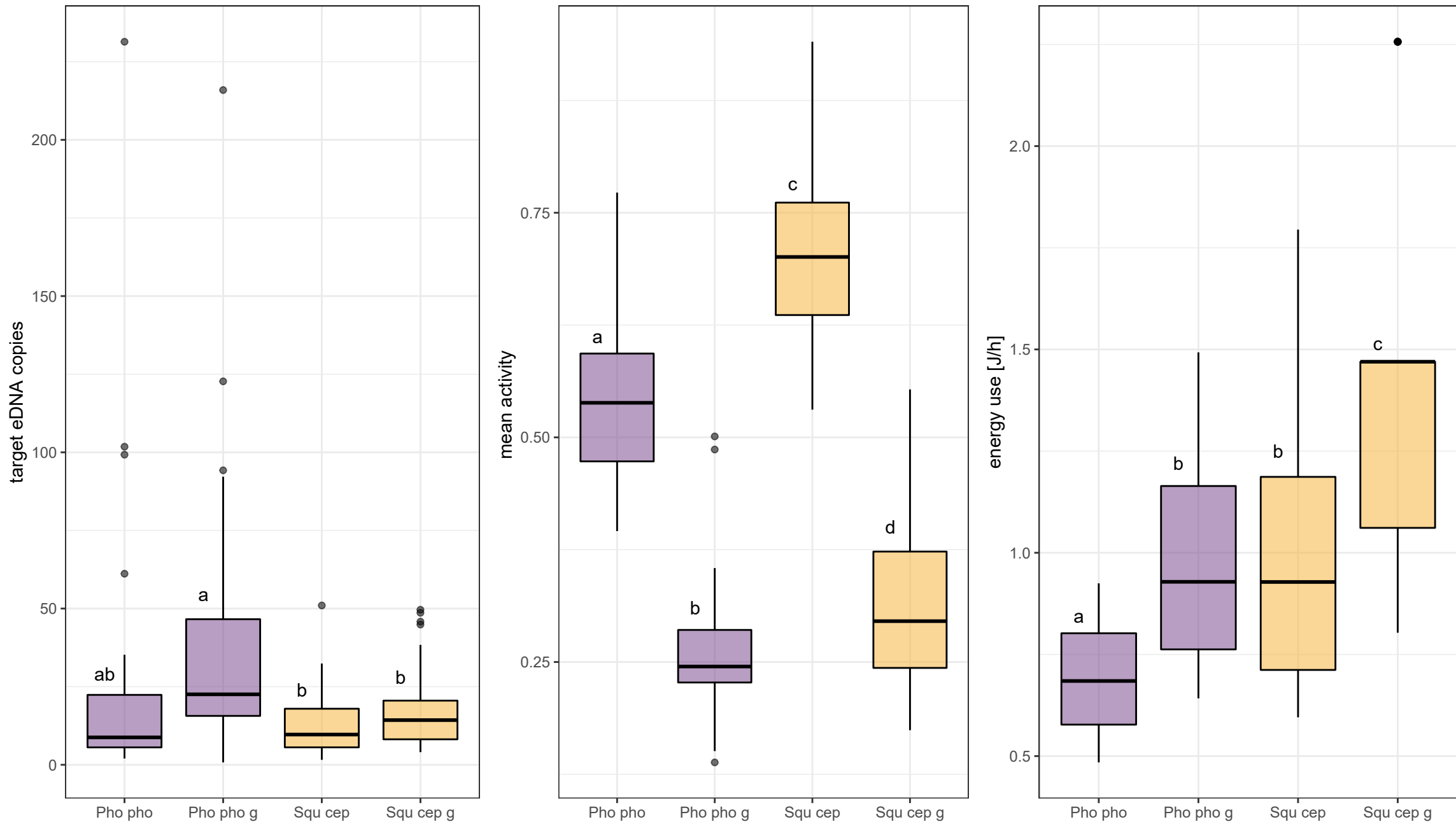


Figure 5

